

CONSERVATION INNOVATION GRANTS

Final Report

Grantee Entity Name: North Carolina State University	
Project Title: Development, evaluation, and demonstration of simple tools for a nitrogen credit trading system involving drainage water management.	
Agreement number: 69-3A75-12-184	
Project Director: Dr. Mohamed A. Youssef	
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Introduction

Agricultural drainage is necessary for crop production on about 25% of cropland in the United States. Improved drainage increases crop yields, but also increases nitrogen (N) loads that can have negative environmental impacts on waters receiving drainage outflows. Drainage Water Management (DWM), also referred to as controlled drainage or managed drainage, is a practice that can reduce N loads while maintaining or improving crop yields. It involves the use of drainage water control structures that are installed at drainage outlets to regulate drainage intensity and outflow according to farming needs. It works by reducing drainage outflow and promoting denitrification reaction to transform dissolved nitrate into dinitrogen gases. Effective management of the structures reduces the amount of water drained and the amount of N lost from the field. The costs of control structures and the time required to effectively manage the structures has limited the acceptance of DWM among growers. An incentive program based on nitrogen trading where farmers could get financial compensation for the effective operation of the drainage water control structure and reducing N loads will likely increase acceptance of DWM and protect the quality of receiving waters. A nitrogen trading program involving DWM will require a simple and accurate tool to calculate the amount of N loss reductions resulting from DWM. This tool will need to consider the many variables affecting N loading from a drained crop field. Also, the input data to this tool must be readily available and verifiable.

Objectives

The objectives of this CIG were to develop and assess a simple tool for quantifying the impacts of DWM on the reduction of drainage flow and N losses from subsurface drained cropland. This tool can be used with an N trading program involving DWM to estimate the annual N reductions resulted from implementing the practice to a specific farm.

Approach

The proposed design of the tool is a set of simple regression equations that can calculate annual reductions in drainage flow and associated N losses for a specific farm as a function of explanatory variables describing essential farm characteristics. The data set used to build the regression equations is generated by thousands of DRAINMOD and DRAINMOD-NII model simulations over representative ranges of soils, climatic conditions, drainage designs, and cropping practices typical for the region of interest. Two separate studies have been conducted: a) Eastern North Carolina; and b) U. S. Midwest. The Midwest includes many states spanning over several climate zones. The states that were chosen for the current analysis were Minnesota, Iowa, and Illinois due to their intensive agricultural presence, their need for water table management, and their spanning of all Midwestern climate zones.

Methods

Development of the Regression Equations

The following steps were followed to develop the regressions equations for the DWM tools for the four states.

- 1) Select locations that represent high crop intensity, different climate zones, and soil types.
- 2) Obtain climate data from National Climatic Data Center of the National Oceanic and Atmospheric Administration (NCDC-NOAA). Thirty years of historical data were obtained.
- 3) Obtain soil data from the Cooperative Soil Survey (MCSS, 2015) and convert to DRAINMOD format with using ROSETTA model (Schaap et al., 2001). Soil series were selected by indicators of poor drainage and crop appropriateness:
 - a) slope <1%;
 - b) depth of soil profile >1 m;
 - c) seasonal water table depth <18 inch,
 - d) a “poorly drained” soil classification by MCSS (2015).
- 4) Use DRAINMOD and DRAINMOD-N II simulations to determine annual drainage volume and annual N load for all combinations of

	<u>Midwest</u>	<u>North Carolina</u>
Location (Climate data for location)	19 counties	4 counties
Soil Series	more than 120	more than 30
Drain spacings	between 9 and 35 m	Between 15 and 55 m
Drain depths	between 70 and 145 cm	between 80 and 150 cm
Fertilizer applications	between 70 to 170 kg N/ha	between 70 and 260 kg N/ha
Surface storage (surface drainage)	1.5 cm, 2.5 cm, 4 cm	0.5 cm, 1.0 cm, 2.0 cm

Simulated scenarios of cropping system and drainage water management include:

Class 1- Continuous Corn under Controlled Drainage;

Class 2- Continuous Corn under Free Drainage;

Class 3- Corn-soybean (Midwest) or Corn-wheat-soybean (NC) rotation under Controlled Drainage;

Class 4- Corn-soybean (Midwest) or Corn-wheat-soybean (NC) rotation under Free Drainage.

- 5) Build regression equation for each scenario using the data sets generated by the DRAINMOD and DRAINMOD-N II simulations. SAS (PROC GLMSELECT)
- a) $Y_d = \beta_1 X_1 + \beta_2 X_2 + \dots + \beta_n X_n + \beta_{C1} X_{C1}$ is the generic equation
- i) The β s represent coefficients estimated by SAS and the Xs represent variables and their interactions to be used in the given scenario

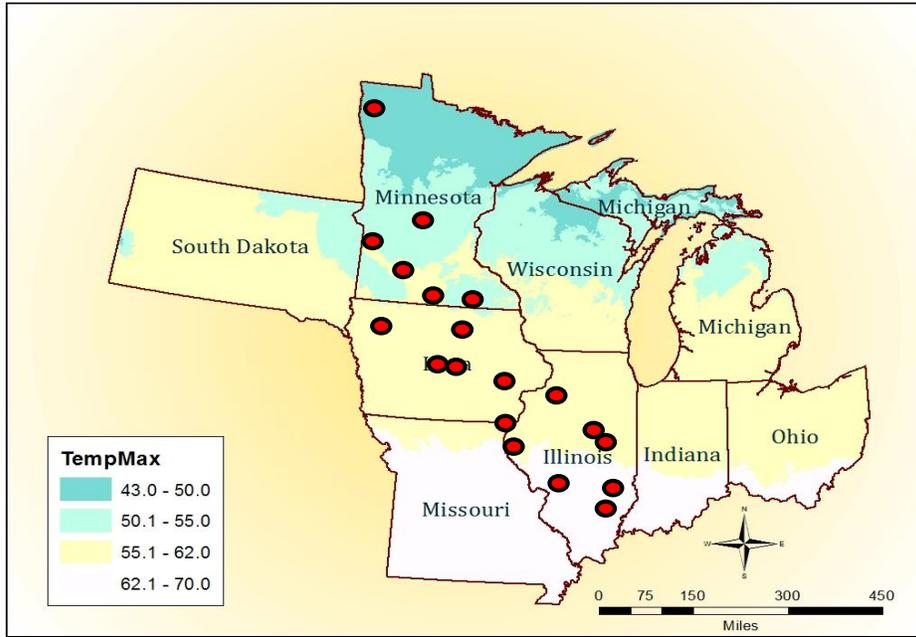


Figure 1. Locations of counties used in the Midwest analyses shown with visual representation of climate divisions: NRCS DATA GATEWAY, 2014

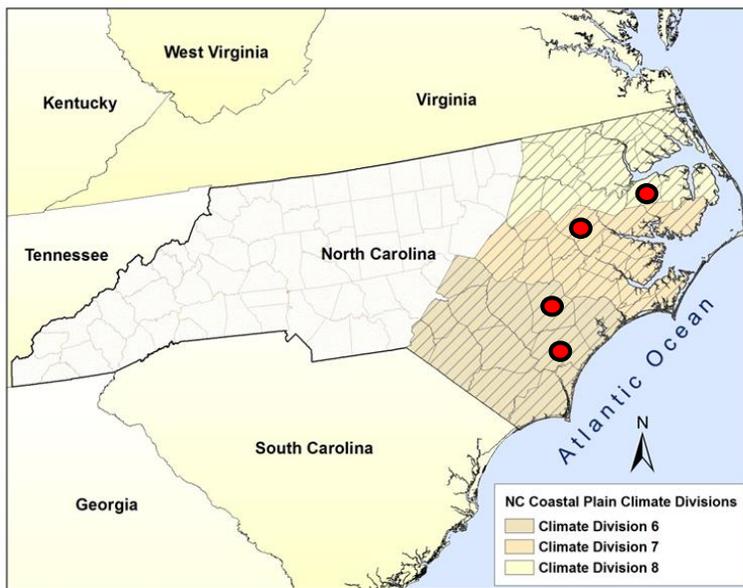


Figure 2. Locations of counties used in the North Carolina analyses shown with visual representation of climate divisions: NRCS DATA GATEWAY, 2014

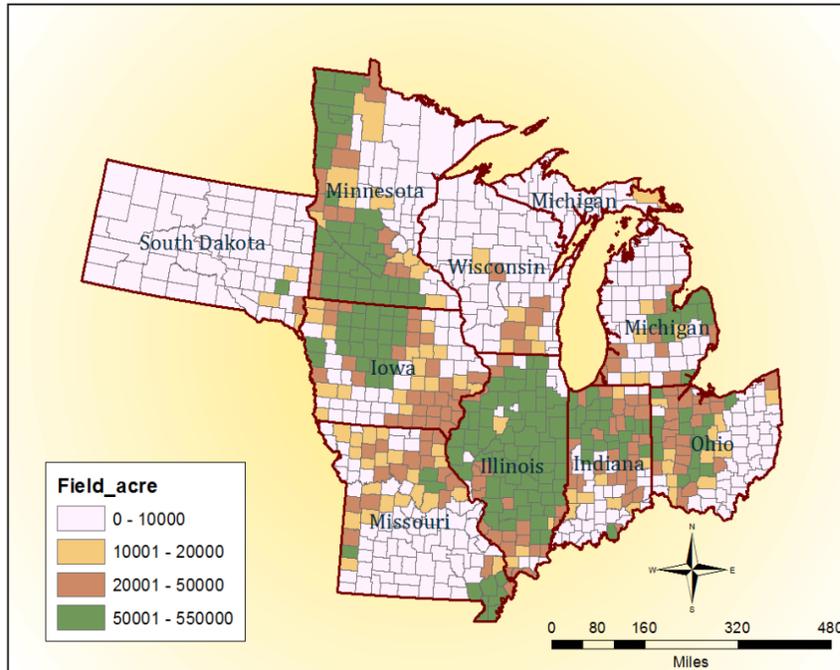


Figure 3. Cropland suitable for drainage water management in the U.S. Midwest. *GIS data sources: <http://datagateway.nrcs.usda.gov> *Cropland county-based tabulated data : <http://www.nrcs.usda.gov/>

The explanatory variables (the inputs) required for the regression equations are listed in the following table.

Explanatory variables for regression equations predicting	
Annual Drainage	Annual Nitrogen Loads
Drain spacing (cm)	Annual drainage either predicted by DRAINMOD or by the drainage regression equation
Drain depth (cm)	Annual relative crop yield (%)
Annual rainfall (cm)	Previous year's relative crop yield (%)
Soil texture (% Sand, % Silt, % Clay)	Fertilizer application rate (Kg N/ha)
Soil Organic Carbon Content (%)	Annual rainfall
Surface storage (Level 1, Level 2, Level 3)	Previous year's rainfall
	Soil Organic Carbon Content (%)
	Ratio of rainfall in growing season to annual rainfall (Midwest only)

Testing of the Regression Equations

The developed regression equations were tested by comparing their predictions to the predictions of DRAINMOD and DRAINMOD-N II model simulations. A detailed description of the approach used for testing the regression equations is given in the manuscript attached to this report.

Results and Deliverables

A) Development and Testing of the DRAINMOD-Based Tools:

1. The outcome of this project is a DRAINMOD-based tool for estimate the impact of drainage water management (DWM) on artificially drained crop lands located in the U.S. Midwest and Eastern North Carolina.
2. The tool comprises a set of regression equations estimating the performance of DWM in terms of reducing drainage flow and NO₃-N losses without the need for using processes based models such as DRAINMOD and DRAINMOD-NII. The regression equations estimate the annual drainage and N loading for four scenarios combining two cropping systems (continuous corn and corn-soybean for US Midwest or corn-wheat-soybean for North Carolina) and two drainage system management: (conventional drainage and controlled drainage).
3. All regression models were developed using the GLMSelect procedure (PROC GLMSELECT) of SAS software (V 9.3, SAS Institute, Cary, NC). This procedure is suitable for models combining both continuous and categorical predictors (More details about the regression analysis and data manipulation are given in the attached manuscript).
4. All regression models require easy-to-obtain parameters that describe local site conditions, including: weather, soil type, drainage system, nutrient management, and crop yields. To use the developed tool, one must select the regression equation that represent the scenario and provide the required input data that represent the local farm conditions.
5. The regression model estimations of annual drainage flow and NO₃-N losses were highly correlated with DRAINMOD simulated values (Figure 4). For the Midwest, adjusted R-squared values were greater than 0.90 for flow and 0.82 for NO₃-N. For NC, adjusted R-squared values were greater than 0.91 for flow and 0.88 for NO₃-N.
6. The developed regression models were compared on a year-by-year basis to the calibrated DRAINMOD and DRAINMOD-NII models for local conditions of an experimental site in eastern NC over 25 years. A similar comparison was conducted for the U.S. Midwest region. The results indicated that the simple regression method provides an adequate alternative to the processes based DRAINMOD suite of models for estimating annual reductions in drainage rates and N mass losses resulting from implementation of DWM. The comparisons between the estimates of the regression models and the predictions of the process based models for a North Carolina farm are presented in the attached manuscript.
7. However, precautions should be considered when using the regression models to estimate DWM performance under extreme weather conditions (very dry or very wet). Under these extreme conditions, the estimates of the regression models tend to considerably deviate from DRAINMOD/DRAINMOD-NII predictions.
8. A Master student and a post-doctoral associate worked on the project. The student is currently writing his thesis and is expected to complete all degree requirements by the end of Fall 2015.

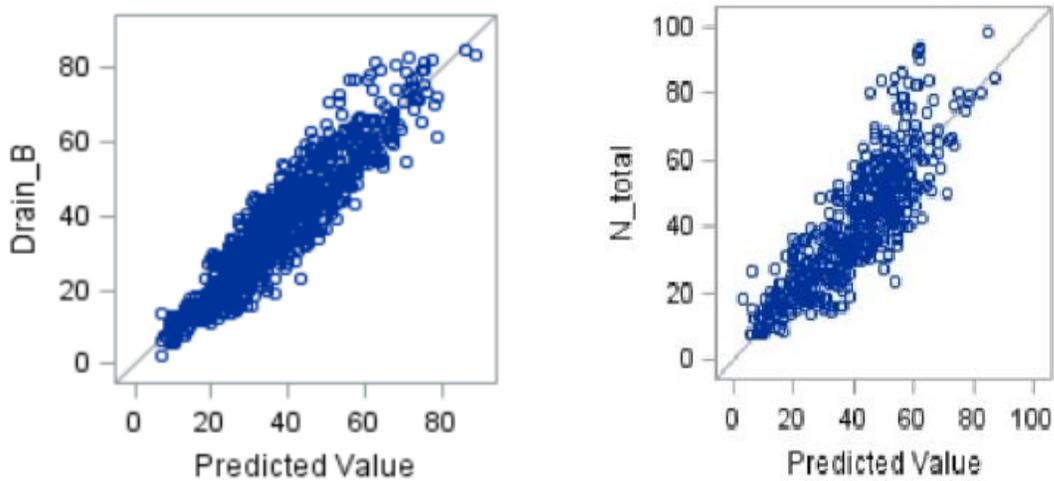


Figure 4. Example output of diagnostic plots comparing DRAINMOD simulation values to the regression predicted values. (PROC REG, SAS procedure).

B) An Online Calculator for the Developed Tools:

An online calculator was designed to help producers and planners to easily apply the tool by simply inserting easy to define parameters describing local site conditions. Screen shots of the online calculator is presented below. This online calculator is primitive and requires significant improvement before it can be readily applied. More funds will be needed to further develop the online interface of the tool.



DRAINMOD-Based Tool for Nitrogen Trading

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DRAINMOD-based Tool for Nitrogen Credit Trading System Involving Drainage Water Management

Nitrogen credit exchange program is a trading market to facilitate pollutant reductions and protect the environment. The current tool provides a simple way to quantify drainage flow and N mass reductions resulting from drainage water management (DWM); an efficient and common conservation practice for drained agricultural lands. Behind the tool are a set of regression model that requires easy to define parameters describe local site environmental and management conditions.

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Drainage Flow

SOIL PROPERTIES		
% Sand Content	% Silt Content	% Clay Content
DRAINAGE DESIGN		
Drain Spacing (m)	Drain Depth (cm)	
Select	Surface Drainage	
CLIMATE		
Annual Precipitation (cm)		
MANAGEMENT		
Select	Crop Rotation	
CALCULATE		

C) Publication and Presentations:

- Negm, L., Youssef, M., Skaggs, R., and Chescheir, G. DRAINMOD-based tools for assessing the impact of drainage water management (DWM) on annual drainage flow and Nitrate losses from cultivated drained soils in Eastern North Carolina. Submitted for The Journal of Agricultural Water Management.
- Negm, L.M., M.A. Youssef, R.W. Skaggs, and G.M. Chescheir, R. O. Evans. 2014. DRAINMOD-based tools for nitrogen credit trading systems involving drainage water management. Presented at ASABE Annual International Meeting, Montreal, Quebec, Canada, July 14-16, 2014.
- Brooks, F.N., L.M. Negm, M.A. Youssef, G.M. Chescheir, and R.W. Skaggs. 2015. Development, evaluation, and demonstration of simple tools for nitrogen credit trading system involving drainage water management. Poster Presentation, 2015 Conservation Innovation Grants Showcase, Greensboro, North Carolina, July 27, 2015.

Program (EQIP) and CIG grant agreement provisions:

1. A listing of EQIP-eligible producers involved in the project, identified by name and social security number or taxpayer identification number; **None**.

2. The dollar amount of any direct or indirect payment made to each individual producer or entity for any structural, vegetative, or management practices. Both biannual and cumulative payment amounts must be submitted. **\$0.0**
3. A self-certification statement indicating that each individual or entity receiving a direct or indirect payment for any structural, vegetative, or management practice through this grant is in compliance with the adjusted gross income (AGI) and highly-erodible lands and wetlands conservation (HEL/WC) compliance provisions of the Farm Bill. **Not applicable.**

1 **DRAINMOD-based tools for quantifying reductions in annual drainage flow and nitrate**
2 **losses resulting from drainage water management on croplands in eastern North Carolina**
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6
7 **Abstract**

8 Nitrogen (N) lechate of drained agriculture has continued to be pervasive in the nation’s water resources.
9 Nitrogen credit exchange program is a trading market to facilitate pollutant reductions and protect the environment.
10 A simple tool suitable for eastern North Carolina (NC) was developed to quantify drainage flow and N mass
11 reductions resulting from drainage water management (DWM); an efficient and common conservation practice for
12 drained agricultural lands. The tool comprises a set of regression equations estimating the performance of DWM as
13 a function of local site conditions. DRAINMOD and DRAINMOD-NII simulations were conducted for a wide range
14 of soil types, weather conditions, and management practices for different locations in eastern North Carolina.
15 Simulation results were used with SAS 9.3 software to develop a set of regression equations to estimate DWM-
16 caused reductions in annual drainage flow and corresponding NO₃-N losses for continuous corn (CC) and corn-
17 wheat-soybean (CWS) cropping systems. The regression model estimations of annual drainage flow were highly
18 correlated with DRAINMOD simulated values with an adjusted coefficient of multiple determination (R^2_{adj}) equaled
19 to 0.91 or higher for different management tiers. Similarly, the regression model estimations of annual nitrate losses
20 achieved an R^2_{adj} of 0.88 or higher for all management tiers. The developed regression models were further
21 compared on a year-by-year basis to the calibrated DRAINMOD and DRAINMOD-NII for local conditions of an
22 experimental site in eastern NC over 25 years. Estimated annual drainage flow and NO₃-N losses were in good
23 agreement with corresponding values simulated by DRAINMOD-based models for CC and CWS under managed
24 and unmanaged drainage modes. In terms of DWM-induced annual reductions in drainage flow and N losses,
25 noticeable differences occurred in several years between predictions of DRAINMOD-NII and the regression models.
26 A comparison based on the 5-year moving average of DWM-induced reductions smoothed out the extreme year-to-
27 year variations and indicated a very similar reduction trends provided by both methods. The results presented in this
28 case study indicated that the simple regression method provides an adequate alternative to the processes based
29 DRAINMOD suite of models for estimating annual reductions in drainage rates and N mass losses resulting from
30 implementation of DWM. Similar tools can be developed for other regions in the US and abroad credit trading
31 system involving DWM.

32
33 **Keywords:** *DRAINMOD, DRAINMOD-NII, regression analysis, agricultural drainage, nitrogen trading,*
34 *controlled drainage.*

36 **1. Introduction**

37 Nitrogen (N) leachate from drained agricultural lands is a pervasive problem to the nation's water resources. The
38 excess of N fertilizer applied on corn fields of the US Midwest have been identified as the main contributor to
39 seasonal hypoxia in the northern Gulf of Mexico, the nation's largest and most productive fishery (Petrolia and
40 Gowda, 2006; Mitsch et al., 2001; Dinnes et al. 2002, Alexander et al., 2008). In the Atlantic coastal plain, excessive
41 levels of nutrients originated from drained agricultural fields have been identified as being a primary cause of
42 eutrophication in the lower Neuse River basin, Pamlico River estuary, and other waters in eastern North Carolina.
43 "The minimization of these N losses is desirable from an environmental standpoint, and a recent interest in
44 discounted reductions of agricultural N losses that might apply to a project downstream from an agricultural area has
45 resulted in the concept of N credits and associated N trading." (Delgado et al., 2008).

46 Controlled drainage (CD), also referred to as drainage water management (DWM), is a practice for reducing NO₃-N
47 losses via drainage water (Skaggs et al., 1994; Evans et al., 1995; Drury et al, 1996; Wesström et al., 2001; Zebarth
48 et al., 2009). Unlike conventional or free drainage (FD) characterized by open drain outlets throughout the year;
49 DWM employs a drainage water control structure at drain outlets to regulate drain flow according to the need for
50 drainage for field operations and crop growth. Previous field experiments that have conducted side-by-side
51 comparisons between managed and unmanaged drainage systems, documented reductions in annual drain flow and
52 N drainage loss resulting from DWM by 20% to 60% (Evans et al., 1995; Pitts et al., 2004; Fausey, 2005; Ayars et
53 al., 2006; Zebarth et al., 2009; Poole et al., 2013, Fang et al., 2012). Managed systems conserve more groundwater
54 and nutrients for plant uptake, especially during relatively dry growing seasons. Ghane et al. (2012) reported that the
55 DWM provided up to 3.3% greater yield compared to FD. A North Carolina study by Poole et al. (2013) indicated
56 that DWM achieved a 10% yield benefit for corn and soybean compared to FD. Drainage water management
57 however, had no significant effect on winter wheat yields due to the wet and cold environment although it is the
58 period when water quality benefits are mostly achieved. Therefore, in many locations where DWM can be
59 effectively applied to improve water quality, the relatively modest yield benefits of the practice are not sufficient to
60 increase nationwide adoption of the practice.

61 Water quality credit trading markets have recently been proposed in the U.S. offering an innovative approach to
62 achieve water quality objectives by controlling pollutant discharge from multiple sources; point and non-point (US
63 Environmental Protection Agency (USEPA), 2007). Through these markets, farmers or stakeholders get monetary
64 compensation based on their application of conservation practices that reduce N losses. Such an approach would be
65 exceptionally effective in promoting the large scale adoption and use of DWM for substantial environmental
66 benefits, as well as, bringing economic and social advantages to farmers and other stakeholders (Corrales et al.;
67 2013 and 2014). One of the largest water quality trading initiatives in the United States is The Great Miami River
68 Watershed Trading Program in Ohio. This program aims to accelerate reductions of phosphorus and nitrogen runoff
69 coming from fields receiving fertilizer and manure applications (Newburn and Woodward, 2012). According to
70 2014 trading program fact sheet released by Miami Conservancy District (MCD), more than 1.14 million credits
71 have been generated since the start of the project in early 2005. This is reported to translate to a 572 ton reduction in
72 nutrients discharge at a cost of 1.6 million dollars paid to agricultural producers for these credits.

73 An essential component of any water quality credit trading system involving DWM is a tool or a method for
74 determining annual N mass conserved by implementing DWM for a specific site and farm conditions (weather, soil,
75 drainage system, cropping system, and farming practices). Several computer simulation models have been
76 developed to predict nutrient losses from agricultural fields as affected by environmental conditions and
77 management practices. Over the past two decades, the DRAINMOD suite of models have been the models of choice
78 for simulating the hydrology and water quality in drained agricultural lands and have been extensively tested and
79 applied nationwide (e.g. Skaggs, 1982; Fouss et al., 1987; Workman and Skaggs, 1989; Cox et al., 1994; Wang et
80 al., 2006a; Ale et al., 2009; Youssef et al., 2006; David et al., 2009; Thorp et al., 2009; Luo et al., 2010; Ale et al.,
81 2012). However, the use of such process-based models requires detailed input data and modeling experience which
82 limit their applicability for this purpose on a production scale. Developing simple decision support tools based on
83 process-based models that can be used by farmers and policy makers for the purpose of nutrient credit trading
84 markets has been well introduced (Delgado et al., 2008; Gross et al., 2008; Corrales et al., 2014; Saleh et al., 2011).
85 Skaggs et al. (2012a) demonstrated the potential of DRAINMOD-based models to formulate trading tools necessary
86 to assess the environmental benefits of installing DWM. Their proposed tool to estimate the effect of DWM on N
87 losses was presented by a simple linear function of annual subsurface drainage and long-term average flow weighted

88 nitrogen concentration. However, the tool oversight temporal and spatial variations in environmental condition and
89 management design which by its turn constrain the tool application to selected configuration of local site conditions.
90 Further research and analysis, utilizing DRAINMOD-based models, can broaden the application of such tools; not
91 only with water quality trading markets, but also with different water resources management studies and
92 applications. For example, this tool can be extremely useful for management of watersheds dominated with drained
93 agriculture.

94 **2. Purpose and Scope:**

95 The purpose of this project is to develop and evaluate a simple tool for quantifying the impacts of DWM on the
96 reduction of drainage flow and NO₃-N losses from subsurface drained cropland in eastern North Carolina. This is
97 the region where DWM was initiated in the mid 1980's for the purpose of reducing nutrient losses. This tool would
98 be essential for any water quality credit trading system that involves the use of DWM. A similar approach will be
99 used to develop tools for crop production systems on drained lands in the U.S. Midwest.

100 Three essential features should be considered in developing this tool: 1) accuracy, 2) simplicity, and 3) data
101 availability. Accuracy is at the core of any water quality credit trading system. Estimates of the credits that all
102 stakeholders can accept is a challenge for the non-point source component(s) of the system. Simplicity is essential
103 since the easy use of the tool will promote its adoption. Data availability is also important since the inputs to the tool
104 must be easy to obtain, standardize, and interpret. The conceptual design of the tool is a set of simple regression
105 equations that can assess annual reductions in drainage flow and associated NO₃-N losses for the region of interest
106 as a function of explanatory variables describing essential farm characteristics.

107 A representative dataset is required to build the proposed regression model and it should comprise a number of
108 observations that consider the inclusion of critical explanatory variables, the temporal and/or spatial variations in
109 each variable, and the response (dependent) variables under investigation (i.e., drainage flow and NO₃-N losses). It
110 is practically impossible, however, to conduct a large enough number of field experiments to quantify DWM
111 performance in response to the wide variation in soil, environmental conditions, and management practices for a
112 given region. Instead, the DRAINMOD suite of models was used to simulate drainage flows and N leaching losses

113 for the wide range of climate year, soil type, drainage design, and farming practices of the NC Coastal plain for
114 both managed (DWM) and unmanaged (conventional drainage) scenarios.

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116 **3. Materials and methods**

117 **3.1 Description of Study Area**

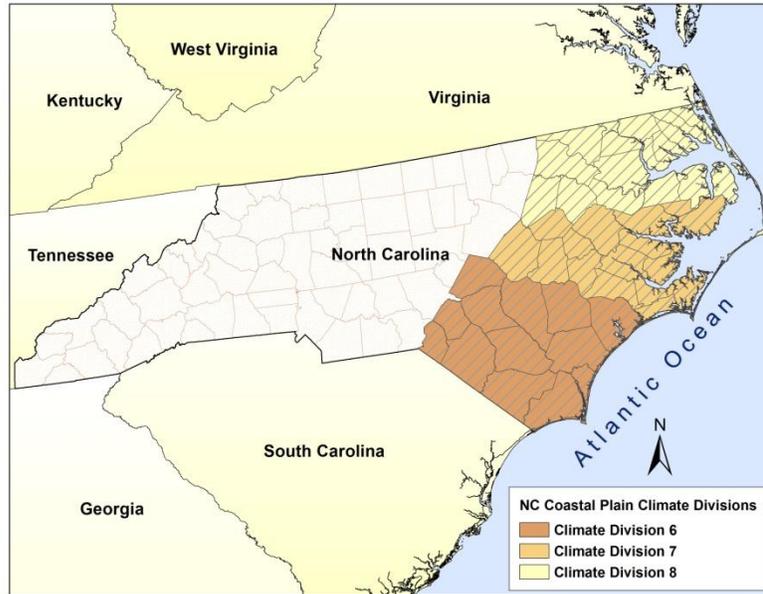
118 The North Carolina Coastal plain encompasses the easternmost counties of the state on a total area of 22,500 square
119 miles that lies between 33° 52' and 36° 33' north latitudes and between 75° and 78 30° west longitudes (Figure 1).
120 This region is characterized by low relief topography becoming more flat and swampy with a larger percentage of
121 the farms artificially drained towards the Atlantic coastal shoreline (Figure 2). The Coastal Plain is the major area
122 for agricultural crop production in the state. The most commonly grown crops in the area are corn, soybean, wheat,
123 and cotton. Installing drainage systems is a common practice by farmers to improve field trafficability and to
124 increase yields and profitability. On average, 40% of the agricultural fields in the area are artificially drained
125 (Pavelis, 1987; Jaynes and James, 2007). Summer temperatures are moderate with daytime high temperatures
126 averaged at 32 °C (90 °F), while winter temperatures are generally mild with 13.4 °C (56.2 °F) average daytime
127 temperatures in the month of January; the coldest month of the season (State Climate office of North Carolina,
128 2013). The annual precipitation in eastern North Carolina averages at 130 cm (51 inches); mostly in the form of
129 rainfall with no distinct dry or wet seasons within a year. Although NC coastal plain is classified as a humid-
130 subtropical climate division, sever dry conditions occasionally occur resulting in substantial losses in the crop yields.

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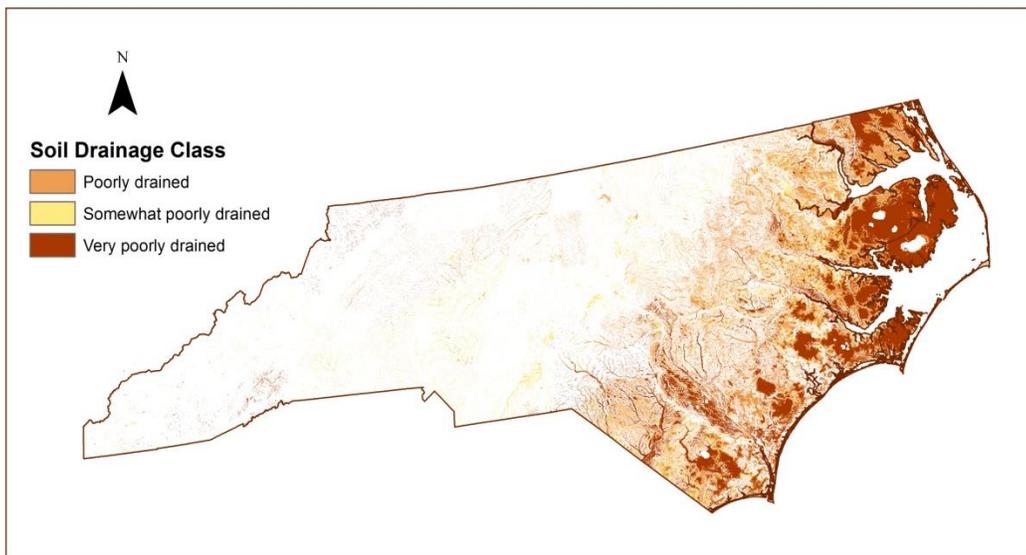
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Figure 1. Map of North Carolina Coastal Plain and Climate Divisions of the Region. (GIS data source: <http://datagateway.nrcs.usda.gov/>)



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Figure 2. A map illustrating the distribution of poorly drained soils in North Carolina. (GIS data source: <http://datagateway.nrcs.usda.gov/>)

143 **3.2 DRAINMOD-based Models description:**

144 **DRAINMOD (Skaggs, 1978)** is a field-scale hydrologic model developed to simulate the hydrology of poorly
145 drained high water table soils. It conducts a one-dimensional soil water balance and predicts subsurface drainage,
146 surface runoff, infiltration, deep seepage, water table depth, and evapotranspiration as affected by changes in
147 weather conditions, crop cover, soil type, and drainage system design. The model applies the Green-Ampt equation
148 to estimate infiltration rates. Subsurface drainage is calculated using Hooghoudt's equation provided that water
149 table is below the surface, otherwise DRAINMOD uses Kirkham's equations when the surface is ponded. Surface
150 runoff is calculated when a user-defined depressional storage is filled as the difference between precipitation and
151 infiltration rates. Daily potential ET (PET) may be internally computed by the model using the Thornthwaite
152 method with monthly correction factors, or daily PET values may be determined independently outside the model
153 by any method (depending on weather data availability) and read in by the model as input data. The model has
154 been widely used to study the effects of drainage design and management on crop yields (e.g. Evans et al., 1991;
155 Wang et al., 2006a; Throp et al., 2009), erosion (e.g. Saleh, 1994), hydrology of high water table soils (e.g. Fouss et
156 al., 1987; Skaggs et al., 1981; Wang et al., 2006a; Throp et al., 2009; Luo et al., 2010), and wetland hydrology (e.g.
157 Skaggs et al., 2005; Jia and Luo, 2006).

158 **DRAINMOD-NII (Youssef et al., 2005)**; a companion model to DRAINMOD, is a process-based model that
159 simulates Carbon (C) and N dynamics of drained cropland. DRAINMOD-NII simulates a detailed N cycle that
160 considers both mineral N (nitrate and ammoniacal forms) and organic N (ON) and their interaction as affected by C
161 cycling. The organic carbon (OC) dynamics is simulated using a C-cycle adapted from the CENTURY model
162 (Parton et al., 1993) by which the soil organic matter is divided into three pools (active, slow, and passive), two
163 above-and below-ground residue pools (metabolic and structural), and a surface microbial pool. The model performs
164 a numerical solution of the multiphase form of the one dimensional advection–dispersion-reaction equation using
165 the finite difference method to simulate N reactive transport in the soil column among three soil N pools: nitrate
166 (NO₃)-N, ammoniacal (NH₃/4)-N, and organic nitrogen (ON). Nitrogen processes simulated by the model include
167 atmospheric deposition, plant uptake, N fixation by legumes, mineralization/immobilization, nitrification,
168 denitrification, ammonia volatilization, and N losses via subsurface drainage, vertical seepage and surface runoff.
169 The model has been tested against field measurements and found to reliably predict N losses in drainage water for a

170 range of soils and locations across the U.S. (North Carolina: Youssef et al., 2006; Iowa: Thorp et al., 2009; Illinois:
171 David et al., 2009; Minnesota: Luo et al., 2010; Indiana: Ale et al., 2012). DRAINMOD-NII has also been tested in
172 Europe (Germany: Bechtold et al., 2007 and Sweden: Salazar et al., 2009).

173 **3.3 Data Collection:**

174 This section describes data collection and processing required to conduct DRAINMOD and DRAINMOD-NII
175 simulations of the hydrology and nitrogen dynamics for a wide range of possible scenarios of local site conditions.
176 DRAINMOD and DRAINMOD-NII simulations of annual drainage flow and annual N losses, respectively, were
177 used to extrapolate the results of field experiments to build the regression model since it is practically impossible to
178 conduct large enough number of experiments.

179 *3.3.1 Weather Data*

180 Historical weather data including hourly precipitations and daily maximum and minimum temperatures were
181 obtained from the North Carolina State Climate Office for different spatially distributed locations across the Coastal
182 Plain. These data covered up to 50 year time span of weather records. A simple analysis using a paired T-test
183 showed that annual rainfall patterns measured for stations within the same climate region were not significantly
184 different ($Pr \geq 0.16$). Therefore, and in order to avoid weather related factors overwhelming the influence of other
185 factors, weather station selections was limited to a few stations at different locations within the North (climate
186 region 8), Middle (climate region 7), and South (climate region 6) of the coastal plain (Figure 1). These weather
187 stations coinciding with four counties; namely from north to south: Washington, Pitt, Duplin, and Pender counties.
188 Rainfall and temperature data were the only climatic data required since Potential evapotranspiration (PET) was
189 computed by the model using the temperature-based Thornthwaite method, with monthly correction factors. All
190 required input weather parameters were then processed to conform to the required DRAINMOD formatting.

191 *3.3.2 Soil Data*

192 The boundary of each of the counties listed above was considered as a buffer zone of the weather station each
193 encompasses; and the soils within each buffer zone and suitable for installing DWM systems were identified through
194 the National Cooperative Soil Survey (MCSS, 2010) and considered in the current analysis as distinct farm units.

195 This selection criterion granted the inclusion of more than 30 different soil series that collectively occupies at least
196 50% of the cropped land suitable for DWM in eastern North Carolina. For many of the selected soil series, soil
197 water characteristics data (soil water content versus pressure head) were available based on previous laboratory
198 measurements on soil cores using standard pressure plate methods. When measured soil water characteristics were
199 not available, soil physical properties were obtained from SSURGO soil data base (MCSS, 2010) and utilized in the
200 ROSETTA model (Schaap et al., 2001) to estimate van Genuchten water retention and unsaturated hydraulic
201 conductivity parameters. Estimated soil water characteristics were further processed using DRAINMOD 6.1 soil
202 utility program to estimate detailed soil related parameters including the volume drained and upward-flux
203 relationships and Green and Ampt infiltration parameters, all as a function of water table depth. Other
204 biogeochemical parameters required to simulate soil carbon and nitrogen dynamics were obtained from previous
205 DRAINMOD-NII applications for NC conditions (Youssef et al, 2006). Organic soils were excluded from this
206 analysis.

207 *3.3.3 Crop management and drainage design:*

208 Two crop rotations were considered in the current study: continuous corn (CC) and corn-winter wheat-soybean
209 rotation (CWS). Crop planting and harvesting dates (Table 1) were defined according to usual operational dates
210 stated by the USDA National Agricultural Statistics Service (NASS) (USDA, 2010). Fertilizer management was set
211 based on recommended practices documented by the Agronomic Division of the North Carolina Department of
212 Agriculture & Consumer Services (NCDA&CS). We considered the reported varying rates of fertilizer application
213 to address all possible levels considered within the region. Fertilized crops; corn and wheat, received 30% of the
214 total amount at planting as a starter, with the remainder sidedressed at the recommended time after planting (Table
215 1).

216 Similarly, drainage system design parameters included in the simulations were set according to common practices in
217 the region to represent realistic conditions for the proposed water management scenarios. Drain depth and spacing
218 were allowed to vary within a range of values listed in the Field Drainage Guide for Coastal Plain Area of North
219 Carolina (NC USDA SCS, 1976). Recommended drain spacings varied between 15 to 50 m placed at depths
220 ranging from 90 to 150 cm. Surface drainage was considered at three different levels: good (depressional storage

221 (DS) ≤ 0.75 cm), fair ($0.75 \text{ cm} < \text{DS} \leq 1.5 \text{ cm}$), and poor ($\text{DS} \geq 1.5 \text{ cm}$). Weir settings that define the control
 222 structure depths under the DWM scenario were adjusted based on the recommendations of the NC Agricultural Cost
 223 Share Program (ACSP).

224 **Table 1: Summary of crop and fertilization management inputs for continuous corn and corn-wheat-soybean**
 225 **rotation at Eastern North Carolina**

Cropping System	Crop	Planting	Harvest	Fertilizer application	
		Date	Date	DAP ¹	Kg N/ha
Continuous Corn	Corn	15 - April	10 - September	0	40 - 66
				30	90 - 154
Corn-wheat-soybean	Corn	15 - April	10 - September	0	40 - 66
				30	90 - 154
	Wheat	10 - November	15 - June	0	27 - 51
				100	63 - 119
	Soybean	19 - June	10 - November	NA ²	NA

226 ¹ DAP= days after planting.

227 ² NA = not applicable

228

229 **3.4 Set-up of hydrologic and water quality simulations:**

230 DRAINMOD and DRAINMOD-NII simulations were conducted to predict annual drainage flow and associated
 231 nitrogen losses for a wide range of environmental factors and management practices within each geographic region
 232 considered. More than ten thousands of simulations were conducted. Input data at the county level were organized
 233 in ASCII-formatted files that list benchmark soils and related soil input files, common drain depths and spacing, as
 234 well as, range variations in fertilization rate. DRAINMOD and DRAINMOD-NII models were integrated through a
 235 batch mode that allows for: 1) updating the models input files for different subsets of scenarios each representing a
 236 unique combination of soil type, drainage design, and fertilizer rate, 2) executing DRAINMOD model to simulate
 237 the hydrology and DRAINMOD-NII to continuously simulate nitrogen dynamics for both managed (DWM) and
 238 unmanaged (conventional drainage) conditions for 20 to 50 years of available historical weather data and for each
 239 proposed scenario, and 3) exporting simulation output to databases that comprise records of simulated annual
 240 drainage flow and N losses (response variables) along with controlling input parameters (explanatory variables)

241 required to build the regression models. The identification of input variables that simulated drainage flow and nitrate
242 losses are sensitive to, was based on previous experimental work and modeling studies (Kladivko et al, 2004 and
243 2005; Youssef et al., 2006; Negm et al., 2014; Skaggs et al., 2005, 2013).

244 The output databases were screened to identify inconsistencies, then merged and organized under two main
245 categories: Database 1 comprised a total of 80,000 records for building a regression model to estimate drainage flow
246 and, Database 2 comprised of more than 200,000 records to estimate N-losses. Database 2 included a greater number
247 of records due to the inclusion of additional parameters that have significant impact on N-losses (e.g.: fertilization
248 rate) while having no impact on predicted drainage flow. Note that estimated relative yield (predicted crop yield/
249 potential yield) predicted by DRAINMOD was considered among the explanatory factors affecting N-leaching
250 losses since it is used by DRAINMOD-NII to simulate N-plant uptake.

251 **3.5 General Linear Regression (GLR) Analysis:**

252 General linear regression (GLR) is an approach by which a response (dependent) variable is related to multiple
253 explanatory (predictor) variables through a linear relationship. However, building a regression model that comprises
254 a relatively large number of predictor variables and based on large number of records is challenging and requires
255 advanced computational techniques. At this phase of the project, we conducted statistical analysis through the SAS
256 9.3 (SAS Institute, Cary NC); a high performance computing software capable of handling large databases and
257 performing robust regression analyses.

258 We chose the GlmSelect procedure “PROC GLMSELECT” to build the proposed regression model since it supports
259 features that are not supported by other regression procedures implemented in SAS software. GlmSelect procedure
260 supports the “Class” statement that is required to distinguish categorical variables from continues variables.

261 Dissimilar to other generalized linear modeling procedures (e.g.: PROC GLM), the GlmSelect includes multiple
262 effect selection methods: forward, backward, stepwise, Least Angle Regression (LAR) (Efron et al., 2004), and
263 Least Absolute Shrinkage and Selection Operator (LASSO) regression (Tibshirani, 1996 and 2011). It allows the
264 modeler to specify different degrees of interaction among different variables. Another option that the GlmSelect
265 procedure supports, utilized herein, is the data partitioning statement that allows for subdividing the input datasets
266 into: a) training data used to build the regression model, b) validation data based on which the predicted error for the

267 model is estimated to avoid regression over-fitting, and c) testing data that played no role in model building and
268 could be used solely to assess the model predictive power. A typical partition fraction assigns 50% for the training
269 data set and 25% each for validation and testing data sets. This procedure was followed in the current study.

270 The regression analysis was initiated by including all candidate explanatory variables (Table 2) that affect a
271 particular constituent (i.e.; annual drainage flow and N losses). Explanatory variables that could be strongly
272 influenced by one or more other explanatory variables were identified through using scatter plots and correlation
273 statistics [p-values and Pearson correlation coefficient (r)]. These variables were excluded from the model to avoid
274 multi-collinearity.

275 The criteria for determining best model from a set of potential models based on the adjusted coefficient of multiple
276 determination (R_{adj}^2). R_{adj}^2 differs from the commonly used coefficient of determination; R^2 , which typically
277 increases as the number of explanatory variables in the regression model increases. R_{adj}^2 compares models that
278 have different numbers of explanatory variables by penalizing models that have additional coefficients (Helsel and
279 Hirsch, 1995) and thereby provides a stronger indicator of the strength of the model.

280 Although the GlmSelect regression, particularly the “Stepwise” selection method, evaluates the contribution of each
281 variable added to or removed from a model, an explanatory variable statistical significance could be exaggerated due
282 to the inclusion of huge number of records used to build the model. Therefore, and through an iterative processes,
283 variables that were identified as significant; i.e., assigned a high t-static, but had no substantial impact on reducing
284 the root mean squared error (RMSE), or on increasing R_{adj}^2 were excluded from the analysis. Since the GlmSelect
285 procedure does not support regression diagnostics, the developed models were further investigated by assessing the
286 relation between simulated values predicted by DRAINMOD-based models with their correspondence estimated by
287 the regression models using a simple linear regression; i.e., PROC REG. Residuals plots and Cook’s distance
288 (Cook’s D) statistic were used to determine outliers or highly influential records.

289

290

291 **Table 2. Candidate explanatory variables for the proposed regression models.**

Category	Explanatory variable	Symbol	Numerical	Categorical	Descriptive statics ¹			Variable Levels ²		
					minimum	maximum	mean	level 1	level 2	level 3
Climate	Annual precipitation, cm	Rain	•		100	160	129.6			
Soil	Silt content, %	Silt	•		5.2	32	18			
	Clay content, %	Clay	•		5	43	20			
	Sand content, %	Sand	•		25	87	62			
	Organic Carbon, %	OC	•		0.4	6	1.3			
	Saturated hydraulic conductivity, cm/hr	Ksat	•							
Sub-surface drainage	Drain depth, cm	D-depth	•		80	150	115			
	Drain spacing, m	Spacing	•		15	55	30			
Surface drainage	Surface drainage conditions	S-Storage		•				good	fair	poor
Crop management & yields	crop relative yield, %	Yield	•		40	100	85			
	crop cover ³	Crop		•				Crn-Wh	Wh-Sb	---
	Fertilizer applied, Kg N/ha/yr	Fert	•		70	260	150			
Preceding year conditions	Climate	L-rain		•				dry	normal	wet
	crop production	L-Yield		•				high	medium	low

292 ¹ Descriptive statics for numerical variables only.

293 ² Different levels for categorical variables only.

294 ³ Used within a corn-wheat-soybean crop management scenario.

295 **4. Results and Discussion**

296 The final product of the current study is a set of regression equations or models designed to estimate annual drainage
297 flow and corresponding annual NO₃-N losses under four tiers of management scenarios: **Tier 1:** A continuous corn
298 cropping system operated under conventional drainage (FDCC), **Tier 2:** A continuous corn cropping system
299 operated under controlled drainage or drainage water management (DWMCC), **Tier 3:** A corn-wheat-soybean
300 cropping system operated under conventional drainage (FDCWS), and **Tier 4:** A corn-wheat-soybean cropping
301 system operated under controlled drainage (DWMCWS).

302 In this section, we report the different regression models, the corresponding fit statistics, and an assessment of their
303 predictive power using the data withheld for model testing. Also, we present an example application that compares
304 the predictions of the regression model to DRAINMOD predictions for a specific site conditions. It should be noted
305 that the regression models were built allowing first degree interaction among different explanatory variables.
306 Allowing a higher degree of interaction did increase model complexity with no additional performance
307 improvements.

308 **4.1 Regression models for annual drainage flow prediction**

309 The selected form of the GLM models developed to estimate annual drainage flow included six numerical variables
310 and two categorical variables. The numerical variables included sand, silt, and clay contents, annual rainfall, and
311 drain depth and spacing. The categorical variables are the surface drainage condition and crop cover in a corn-
312 wheat-soybean rotation tiers. Although DRAINMOD predictions of drainage flow have been found to be sensitive
313 to changes in the soil lateral saturated hydraulic conductivity (K_{sat}) (Hann and Skaggs, 2003; Wang et al., 2006b),
314 the values of K_{sat} were found to be highly correlated with soil texture parameters; sand ($r = 0.67$), silt ($r = -0.49$), and
315 clay ($r = -0.63$). This finding is expected as numerous studies have shown the dependency of K_{sat} on soil texture (e.g.
316 Saxton et al., 1986; Tietje and Hennings, 1996; Rawls et al., 1998; Kosugi, 1999). Therefore, K_{sat} were not included
317 in the regression models. Predicted annual drainage flow was not significantly affected by parameters related to crop
318 management, yields, as well as, parameters describing preceding climatic conditions. Annual precipitation was the
319 most significant parameter controlling the variation in drainage flow with an $r = 0.52$. These results showing the

320 sensitivity of drainage flow to different explanatory variables agreed with previous experimental findings (Youssef
321 et al., 2006; Kladvik et al, 2005; Wang et al., 2006a; Jaynes and Colvin, 2006).

322 A preliminary assessment of the regression models' performance showed that their estimation of annual drainage
323 flows in either extreme dry or extreme wet years were associated with very high residuals and Cook's D values. As a
324 result, records with the most extreme 5% of the rainfall distribution (Rain < 99 cm or Rain > 160 cm) were
325 eliminated from all data subsets and the regression process was repeated. Therefore, the regression equations may
326 not provide accurate estimates of the annual drainage flow if to be applied on years with extremely high or low
327 annual precipitation.

328 The response variable Y_H ; annual drainage flow (cm), can be estimated from the following equation:

$$329 \quad Y_H = \beta_1 X_1 + \beta_2 X_2 + \beta_3 X_3 + \dots + \beta_n X_n +$$
$$330 \quad \beta_{L1} X_{L1} + \beta_{L2} X_{L2} + \dots + \beta_{Lm} X_{Lm} \quad (\text{Eq. 1})$$

331 where m and n are the number of equation terms without and with a categorical variable, respectively. β_i is the
332 regression coefficient corresponds to i^{th} equation term (X) representing a numerical variable or an interaction
333 between two numerical variables for $i = 1, 2, 3, \dots, n$. β_{Lj} is the regression coefficient corresponds to j^{th} equation term
334 (X_{Lj}) representing a categorical variable, or an interaction between a categorical and numerical variable, or an
335 interaction between two categorical variables.

336 All the equation terms and corresponding parameter estimates for different management tiers are listed in Table S1
337 and Table S2. In most cases the selected explanatory variables can be easily defined with no need for intensive or
338 expensive field measurements. This was the goal of the study, to develop a simple tool that can be used with easily
339 accessible data.

340 The four tiers' regression models developed to estimate annual drainage flow achieved an R^2_{adj} higher than 0.88, and
341 a root mean squared error (RMSE) less than 10.5 cm (Table S1 and Table S2). These numbers indicate a high
342 correlation between simulated and estimated annual drainage flow which is mainly attributed to simultaneously
343 assessing the effects of various explanatory variables on the response variable through the utilization of the
344 GlmSelect procedure and through a factorial design structure of the modeled data (Pedhazur, 1997). The factorial

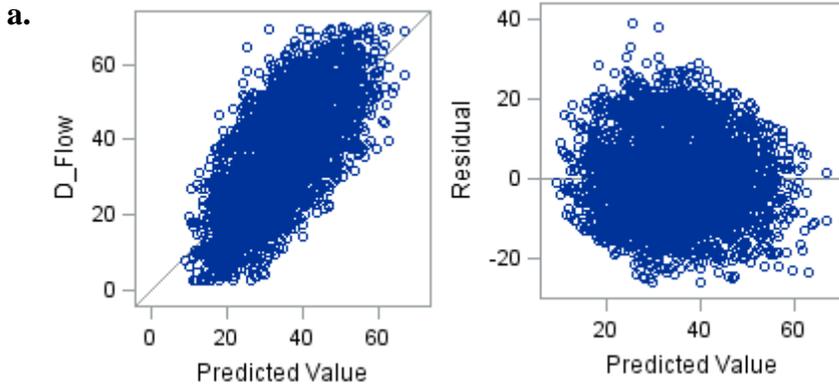
345 design considers a multi-dimensional space that combines all possible combinations of several factors, each with
346 several numerical or categorical levels. However, caution must be used in applying the models for years with
347 extremely low or high annual rainfall.

348 The different terms of the developed equations (Table S1 and S2) emphasized that joint effects between parameters
349 generally govern the variation in the response variable (Y_H) more than what the main effect of each explanatory
350 variable. This could be attributed to the strong inherent interaction among different environmental and management
351 design variables in their contribution to controlling the system performance.

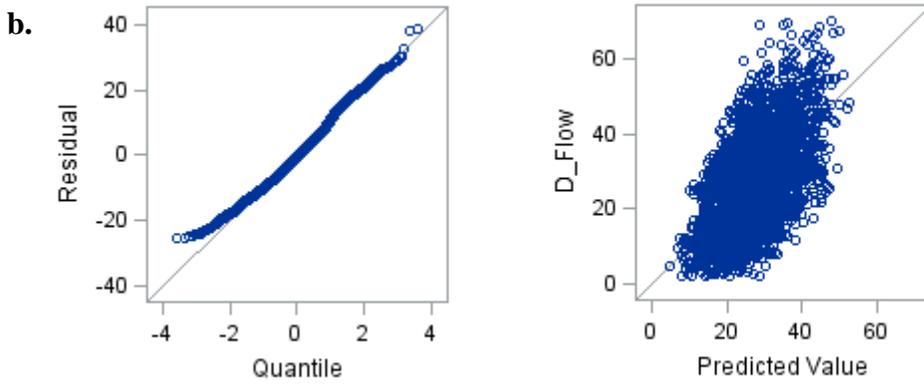
352 Figures 3 illustrates comparison between DRAINMOD simulated annual drainage flow and their corresponding
353 estimated by the four-tiers regression models utilizing the dataset withheld for model testing. Both simulated and
354 estimated annual drainage values were highly correlated with a R_{adj}^2 equals to 0.93, 0.92, 0.94, and 0.91 for tier 1,
355 tier 2, tier 3, and tier 4 regression models, respectively. These values are as good as those estimated for the training
356 dataset indicating an unbiased regression model. The random distribution of the residual in the diagnostic residual
357 plots, as well as the proximity of the quantile plots to straight lines (Figure 3) indicate that regression model
358 assumptions including residuals normality and homoscedasticity (equal variance) were preserved.

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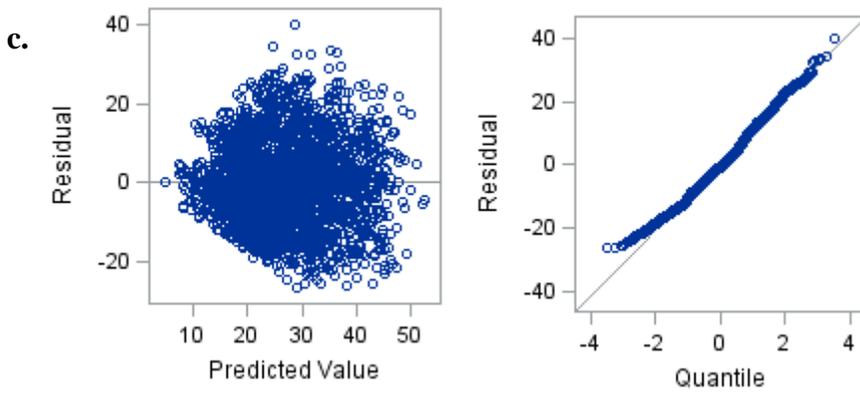
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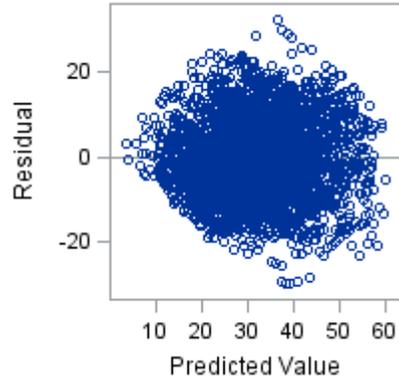
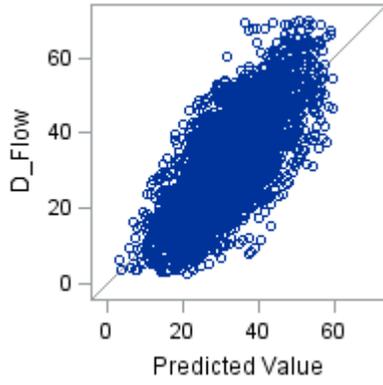
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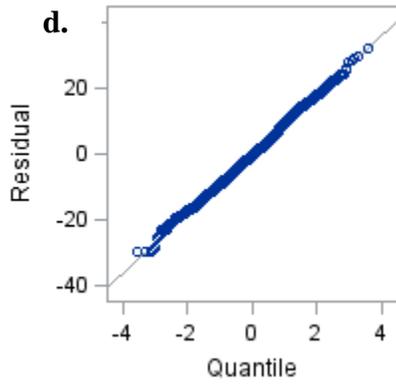
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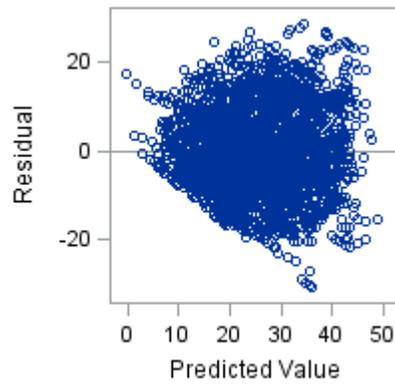
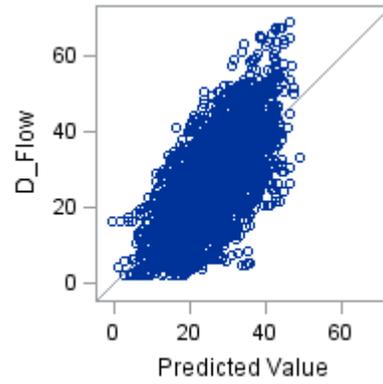
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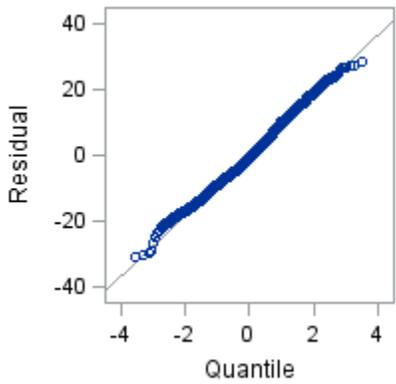
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378 **Figure 3. Correlation between simulated and estimated annual drainage flow (cm/year), scatter residual plots,**
379 **and residual-quantile plots using the data withheld for model testing for a) FDCC, b) DWMCC, c) FDCWS,**
380 **and d) DWMCWS management scenarios.**

381 **4.2 Regression models for predicting annual NO₃-N losses**

382 The strategy used to develop regression models for estimating annual NO₃-N losses was similar to that described in
383 Section 4.1 for drainage outflows. We started building the model by implementing all categorical and numerical
384 variables that showed significant effect on estimating annual drainage flows and then added parameters related to
385 crop management. However, the large number of explanatory variables caused the resulting models to be very
386 complicated. Several modeling and experimental studies documented that rates of NO₃-N leaching losses are highly
387 correlated with rates of drainage flow (Kladvko et al., 2004; Youssef et al., 2006; Ale et al., 2009; Luo et al., 2010;
388 Skaggs et al., 2012a; Negm et al., 2014). Thus, predicted annual drainage flow was used as an explanatory variable
389 to estimate annual NO₃-N loss to simplify the equation. Using a simple correlation statement; i.e. PROC CORR,
390 annual drainage flow rate was the most correlated parameter with annual nitrate losses ($r > 0.5$).

391 The final regression models to estimate annual NO₃-N losses included four numerical variables and two categorical
392 variables. The numerical variables included annual drainage flow, soil organic carbon, relative yield, and fertilizer
393 application rate. Selected categorical variables included preceding year crop production level and climate conditions,
394 in addition to crop cover in corn-wheat-soybean rotation settings.

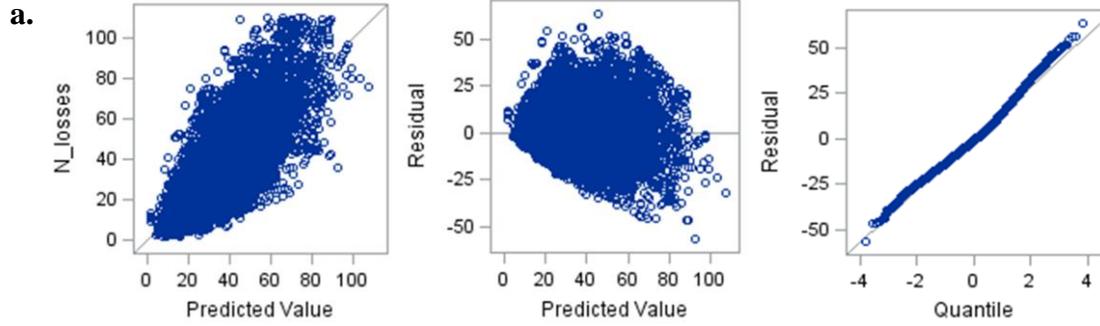
395 The response variable Y_N ; annual NO₃-N losses (Kg N ha⁻¹), can be estimated from the regression model (Equation
396 1) where all the equation terms and corresponding parameter estimates for different management tiers are listed in
397 Table S3 and Table S4. Annual drainage flow may be estimated as described in section 4.1. An alternative is to run
398 DRAINMOD model directly to compute annual drainage flow for a local site conditions of interest. Although this
399 alternative is expected to yield more accurate estimation of drainage flow, it does require modelling experience and
400 data availability.

401 The developed regression models for estimating annual NO₃-N losses resulted in an adjusted R² of 0.88 or higher for
402 all management tiers indicating high correlation between DRAINMOD-NII simulated and estimated annual NO₃-N
403 losses. Incidences at which the estimated annual drainage losses were poorly correlated with DRAINMOD-NII
404 simulated values (high value of the residual) were scattered between the upper and lower bounds of the response
405 variable. Such differences could be attributed to DRAINMOD models capability of simulating dynamic
406 environmental conditions and their interaction with management practices. The regression models are not capable of

407 considering such interactions. For example, the intensity of rainfall events and their timing relative to the growing
408 season and fertilizer application dates in a given year affect N concentration in drainage waters, and, in turn, annual
409 N load. Rainfall at a different intensity and date with respect to fertilization may result in a very different annual N
410 loss, even though the annual rainfall and fertilization are similar.

411 Figure 4 illustrates the diagnostic plots for assessing the ability of the regression models to estimate annual N loads
412 in comparison to DRAINMOD-NII simulated values for the four tiers of management scenarios. The achieved R^2_{adj}
413 equaled to 0.9, 0.88, 0.92, and 0.92 for tier 1, tier 2, tier 3, and tier 4 regression models, respectively. This indicates
414 consistently good performance of the developed regression equations in estimating N drainage losses under both
415 conventional and controlled drainage. This performance was achieved despite lumping a large number of interacting
416 complex hydrological and biogeochemical processes. This simple model provides a calculation tool for estimating
417 the effect of implementing DWM on N loads at the field outlet.

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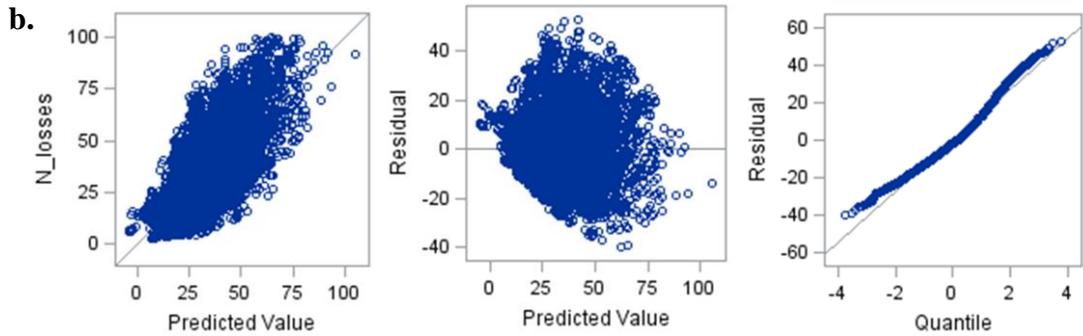
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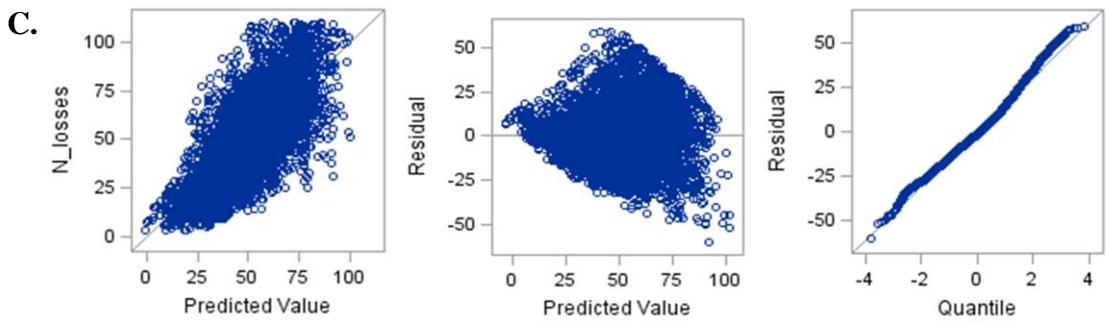
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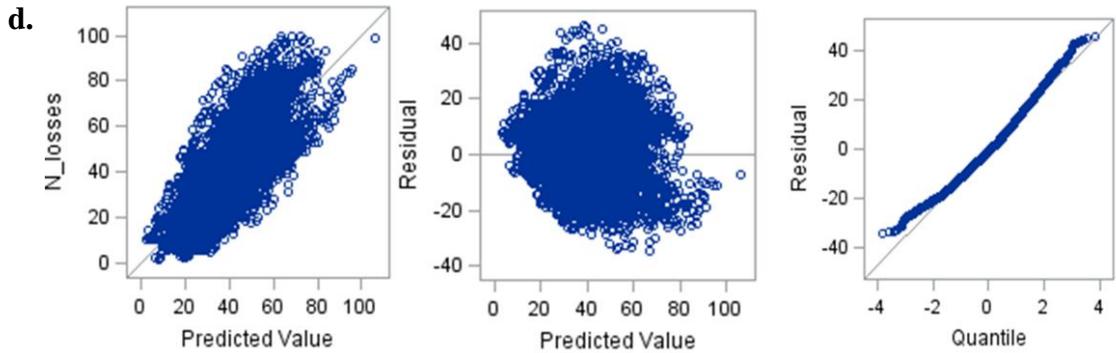
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439 **Figure 4. Correlation between simulated and estimated annual drainage flow (cm/year), scatter residual plots, and**
 440 **residual-quantile plots using the data withheld for model testing for a) FDCC, b) DWMCC, c) FDCWS, and d)**
 441 **DWMCWS management scenarios.**

442

443 4.3 Case Study:

444 In the this section, the regression models' estimation of drainage flow and NO₃-N leaching losses were compared, on a
445 year by year basis, to the predictions of DRAINMOD and DRAINMOD-NII, calibrated for a drainage water management
446 site located at the Tidewater Research Station (TRS) near the town of Plymouth, in the North Carolina lower coastal plain
447 (Skaggs et al., 2012b, Youssef et al., 2006). Throughout this example application, we refer to drainage flow and NO₃-N
448 losses predicted by DRAINMOD and DRAINMOD-NII as “simulated”, while the values calculated by the regression
449 models are referred to as “estimated”.

450 4.3.1 Description of the Agricultural System

451 The soil on the nearly flat, 13.8-ha site is Portsmouth sandy loam, which is very poorly drained under natural conditions.
452 The surface soil horizon characterized by a relatively high organic carbon (3.5%). The subsurface drain pipes were spaced
453 at 23 m and placed at 118 cm below the soil surface. The site is planted to a corn-wheat-soybean rotation; however, a
454 continuous corn scenario has also been proposed for complete demonstration and comparison purposes. A 25-year period
455 (1976-2000) of measured weather data (daily temperatures and precipitation records) was utilized for this example
456 application. Table 3 summarizes variables that describe site characteristics required by the regression models. Detailed
457 description of DRAINMOD and DRAINMOD-NII parameterization is described by Youssef et al. (2006) and Skaggs et al.
458 (2012b).

459 **Table 3. Summary of soil property and site parameter inputs for the TRS site.**

460	Variable	Value
461	Drain depth, m	1.18
462	Drain spacing, m	23.0
463	Depressional Storage, cm ¹	1.25
464	Sand content, %	75.5
465	Silt content, %	14.0
466	Clay content, %	10.5
467	OC, %	3.5
468	Fertilization, Kg N ha ⁻¹	
469	<i>Corn</i>	175
470	<i>Wheat</i>	120
	<i>Soybean</i>	NA

471 ¹ The 1.25 cm of depressional storage is equivalent to Level 2 of surface storage in the regression equation.

472 ² In a corn-wheat-soybean rotation, annual applied fertilizer for a CW year depended on the total amount of fertilizer applied for
473 corn plus the starter amount applied for wheat, and for a WS year, the annual fertilizer applied equaled to the sidedressed
474 application for wheat.

475 **4.3.2 Annual Drainage Flow**

476 Annual drainage flow simulated by DRAINMOD and estimated by the regression models are visually compared in Figure
477 5. For the FDCC and DWMCC, the Percent Difference (PD) between estimated and simulated annual drainage flow was
478 within $\pm 15\%$ in over 56% of the years, and within $\pm 25\%$ in over 82% of the years. Over the 25-year period, the PD
479 between estimated and simulated average annual drainage flow was -2.3% for the FDCC (estimated = 39.8 ± 12.7 cm,
480 simulated = 40.8 ± 13.5 cm), and PD was 1.6% for the DWMCC (estimated = 32.2 ± 12.6 cm, simulated = 31.7 ± 10.6 cm).
481 Years when estimated drainage substantially differs from simulated values were mostly associated with years experienced
482 extreme climate conditions; years with extreme annual precipitation and/or years that has monthly precipitation patterns
483 that sharply deviated from long-term means. For example, in the extreme dry year of 1993 (annual precipitation = 100 cm)
484 estimated annual drainage was 42% and 47% less than their corresponding values simulated by DRAINMOD for the
485 FDCC and DWMCC, respectively. In 1999 when annual perception were similar to long-term averages, estimated and
486 simulated annual drainage flows respectively were 42.6 and 34.3 cm (PD = +24%) for the FDCC, and were 35 and 26.9 cm
487 (PD = +30%) for the DWMCC. These differences could be mainly attributed to the significant divergence of seasonal
488 rainfall from normal conditions. The period of February till August 1999 which received 56.5 cm of rainfall and reported
489 as the third driest period in a 49-year record, was followed by a series of hurricanes and tropical storms with a total of 55.5
490 cm of rainfall during September and October 1999. This was the wettest period in a 49-year historical weather record
491 (1951 -1999) (Shelby et al., 2006). Under such conditions, DRAINMOD simulated near zero drainage in the prolonged dry
492 period followed with large drainage volumes associated with excessive surface runoff in the short but extremely wet
493 period. However, estimated drainage values by the regression model were close to long term estimated drainage values
494 since it is based on annual rainfall amounts and does not consider the seasonal extremes.

495 For the FDCWS and DWMCWS, PD between estimated and simulated annual drainage flow was within $\pm 15\%$ in over
496 52% of the years, and within $\pm 25\%$ in over 76% of the years. Estimated and simulated average annual drainage over the
497 case study period were 40.8 and 38.5 cm for the FDCWS (PD = 5.1%); 34.1 and 31.3 for the DWMCWS (PD = 9.1%).
498 Similar to the continuous corn management scenarios, seasonal distribution of rainfall events was a critical factor
499 impacting the difference between estimated and simulated annual drainage flow in some of the years. For 1999, estimated

500 annual drainage flow were 36% and 19% higher than the simulated values respectively for FDCWS and DWMCWS.
501 Under the extreme dry conditions of 1993, estimated annual drainage flow were 27.7% and 30.7% lower than the simulated
502 values for the FDCWS and DWMCWS, respectively. Another exceptionally dry year was 1997, which was similar to 1993
503 in terms of annual rainfall and crop cover (corn-wheat). The regression model estimated similar annual drainage flows for
504 the two years. DRAINMOD simulated lower drainage flow in 1997 than in 1993. In 1997, around 50% of annual rainfall
505 occurred during the hot growing season when evapotranspiration demands normally exceed drainage rates. In contrast
506 70% of the annual rainfall in 1993 occurred during the fallow period resulting in greater drainage volumes. Such
507 differences cannot be predicted by the simple regression equations. Yet, overall, they do a very acceptable job in
508 predicting annual outflows.

509 A fair correlation between estimated and simulated percent reduction in annual drainage due to DWM can be inferred from
510 the year-by-year comparison illustrated in Figure 6 for CC and CWS. The 5-year moving average curves reduced the
511 effects of temporal variations and identified no over- or under-estimation trends predicted by the regression models, in
512 comparison to annual flow reductions predicted by DRAINMOD.

513 Overall, the simple regression models showed comparable performance to DRAINMOD in estimating annual drainage
514 flow under different management and climate scenarios. They provide a simple, reliable method of estimating annual
515 drainage flow, which can be used to estimate annual $\text{NO}_3\text{-N}$ losses as described in the following section. These equations
516 may also be useful in other water conservation and environmental management applications where estimates of annual
517 drainage outflows are needed.

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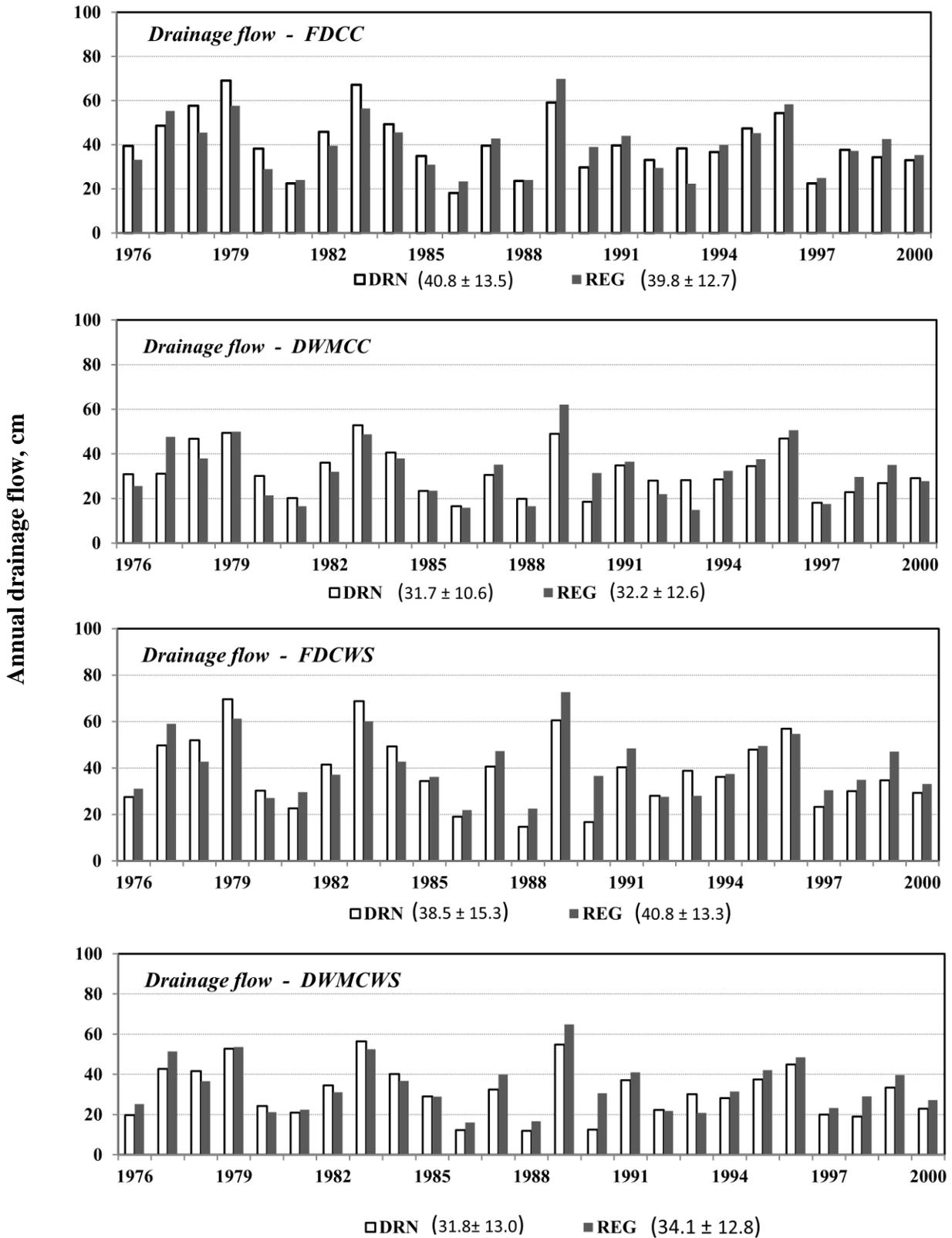
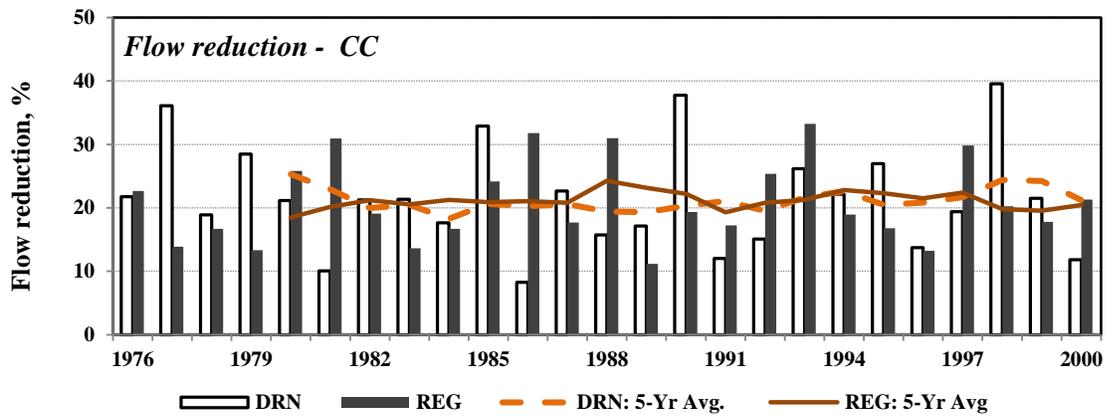
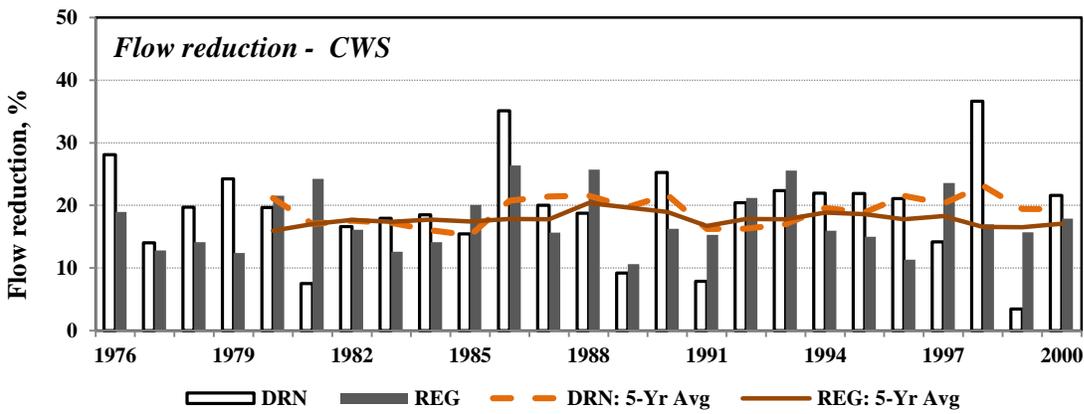


Figure 5. Annual drainage flow as simulated by DRAINMOD (DRN) and estimated by the regression models (REG) for different management tiers. (Values between parentheses represent average annual drainage flow \pm standard deviation, all units are in cm).



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528 Figure 6. Reduction in annual drainage flow (%) due to DWM as simulated by DRAINMOD (DRN) and estimated
 529 by the regression models (REG) for continuous corn (CC) and corm-wheat-soybean (CWS) management tiers. Solid
 530 line represents the 5-year moving average of reduction in annual drainage flow (%) as estimated by the regression
 531 model (REG: 5-Yr Avg), and the dashed line represents the 5-year moving average of reduction in annual drainage
 532 flow (%) as simulated by the DRAINMOD (DRN: 5-Yr Avg).

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542 **4.3.3 Annual Nitrate Losses**

543 Utilizing the estimated annual drainage flow discussed in section 4.3.2; annual NO₃-N losses were estimated for the
544 same management tiers considered in this example application. Figure 7 presents the annual NO₃-N losses estimated
545 by the regression models as compared to corresponding values simulated by DRAINMOD-NII for the FDCC,
546 DWMCC, FDCWS, and DWMCWS. The Percent Difference (PD) between estimated and simulated annual N losses
547 was within ±25% over 76%, 72%, 76%, and 56% of the 25 years for FDCC, DWMCC, FDCWS, and DWMCWS,
548 respectively.

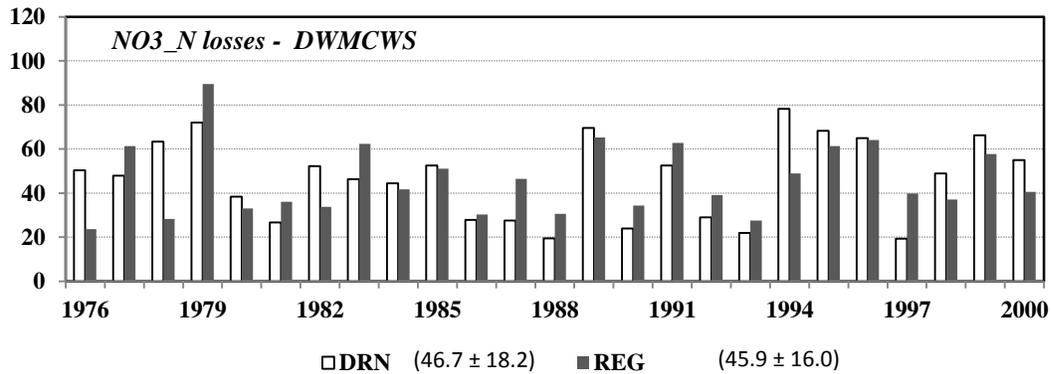
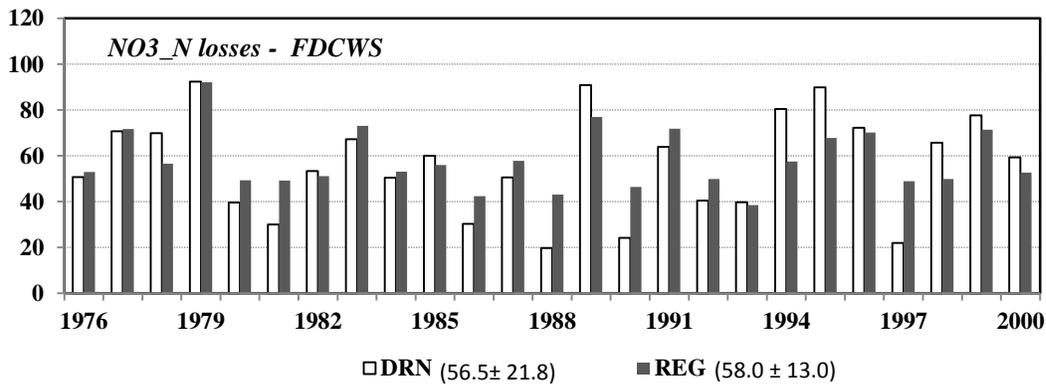
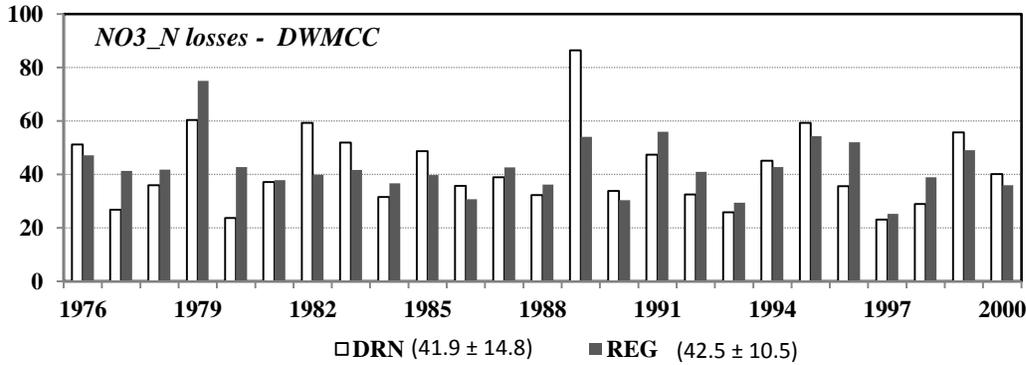
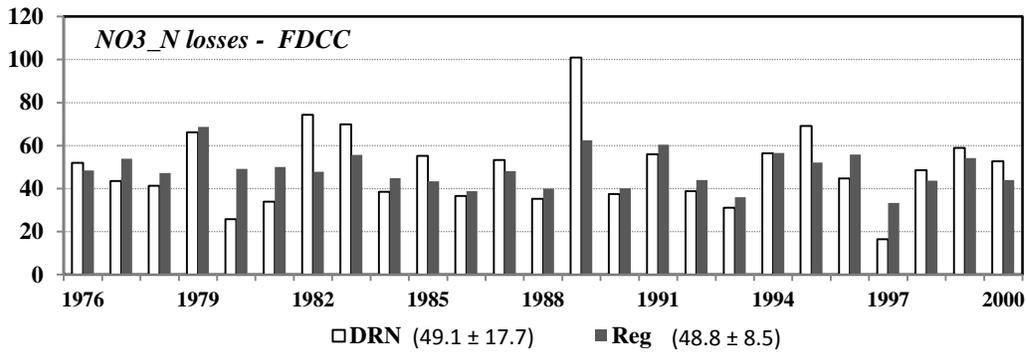
549 Larger PD's occurred in some years for both rotations and drainage modes. As an example, in years 1982, 1983,
550 and 1989; estimated losses were noticeably lower than simulated values for the FDCC and DWMCC (Figure 7).
551 This could be mainly attributed to the timing of heavy rainfall events that coincided or shortly followed site
552 fertilization causing DRAINMOD-NII to simulate high N losses early in the growing season. In addition, estimated
553 drainage flows were lower than simulated values for 1982 and 1983. In the exceptional wet year of 1989 (annual
554 precipitation = 176 cm) heavy rainfall events coincided with or closely followed days when fertilizer was applied.
555 DRAINMOD-NII simulated excessive NO₃-N losses (30 Kg N/ha for FDCC and 23 Kg N/ha for DWMCC) during
556 the very wet month of March as nutrient residues turned over the preceding dry year of 1988 (annual precipitation =
557 103 cm). For the CWS scenarios, estimated and simulated annual NO₃-N losses for 1989 were in close agreement. In
558 this year (1989), the field was planted to corn following soybean, which received no fertilization in 1988.
559 Therefore, nutrient residuals carried over to 1989 was not as significant as in a CC scenario. Again, the timing of
560 rainfall events relative to nutrient management, growing season and crop cover dramatically affects DRAINMOD-
561 NII simulations. The regression models do not consider such factors, resulting in the discrepancies in N losses, and
562 the effect of DWM on N losses predicted by the two approaches.

563 Average annual NO₃-N losses over the 25 years estimated by the regression models agreed well with DRAINMOD-
564 NII predictions for the FDCC (estimated = 48.8 ± 8.5 Kg N/ha; simulated = 49.8 ± 18.5 Kg N/ha), and for the
565 DWMCC (estimated = 42.5 ± 10.4 Kg N/ha; simulated = 41.2 ± 14.4 Kg N/ha). Similarly, estimated average annual
566 NO₃-N losses were in close agreement with simulated values for the FDCWS (estimated = 58.0 ± 13.0 Kg N/ha;
567 simulated = 56.5 ± 21.8 Kg N/ha), and for the DWMCWS (estimated = 45.9 ± 16.0 Kg N/ha; simulated = 46.7 ±
568 18.2 Kg N/ha). The fact that the regression approach predicted results similar to the long-term annual averages
569 predicted by DRAINMOD-NII is expected since a regression model usually performs better as an estimator of
570 normal conditions, than for other conditions that occur less frequently.

571 The potential application of these regression models in N trading markets was the main motive for developing the
572 current regression model as a tool to estimate DWM-induced reductions in N losses. Figure 8 illustrates annual
573 reductions due to DWM for CC and CWS scenarios as predicted by the regression approach versus DRAINMOD-
574 NII simulations. Average annual reduction in NO₃-N losses due to DWM as estimated by the regression model for
575 the CC scenario was 6.3 Kg N/ha versus 8.6 Kg N/ha predicted by DRAINMOD-NII. For the CWS, estimated
576 annual reductions averaged at 6.6 Kg N/ha versus a simulated value of 7.2 Kg N/ha.

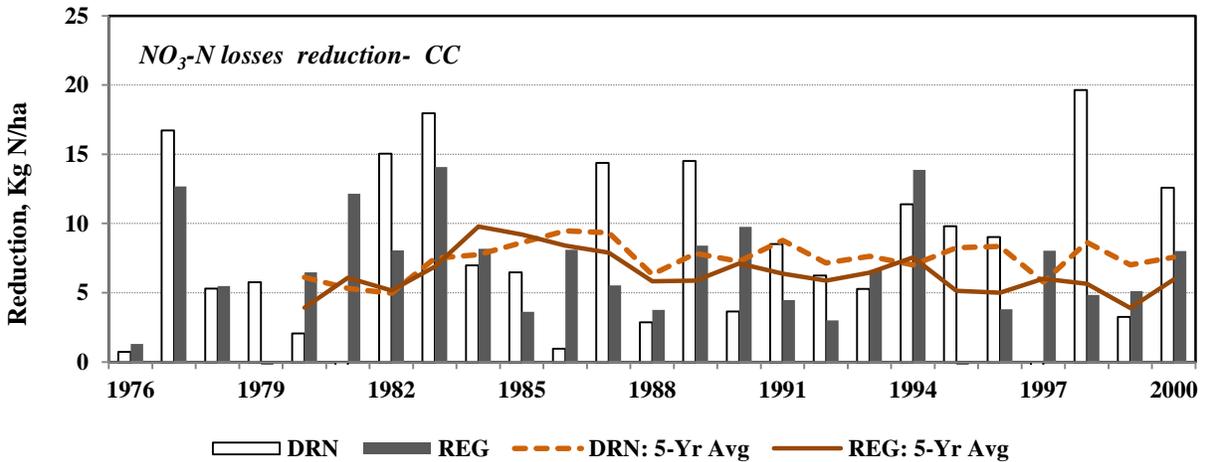
577 On a year-by-year basis, estimated reductions did not agree well with DRAINMOD-NII predictions in many of the
578 years (Figure 8). For CC, noticeable difference between estimated and simulated reductions was in years with
579 extreme dry growing seasons (i.e.; 1981, 1986, and 1997) associated with low crop yields. For the normal-to-wet
580 years followed these dry years (i.e.; 1982, 1987, and 1998), the regression model tends to estimate lower N losses
581 reductions compared to that simulated by DRAINMOD-NII. Estimated reductions in 1979 and 1995 didn't agree
582 with simulated values. These years experienced very wet growing seasons causing significant crop yield reductions
583 (as simulated by DRAINMOD), especially under the controlled drainage mode that caused additional loss to crop
584 yields due to increased wet stresses. For the CWS, discrepancies between estimation of DWM-caused reductions are
585 noticeable in several years. This was especially the case in the first 6 years of simulation when the regression model
586 estimation of annual NO₃-N losses didn't agree well with DRAINMOD-NII simulated values under the controlled
587 drainage mode.

588 There was less variability when the comparison is based on the 5-year moving average of the predicted annual
589 reductions of NO₃-N losses (Figure 8), as overestimation in some years was counterbalanced by underestimation in
590 other years within the 5-year period. For CC the effect of DWM on the 5-year moving average of N losses as
591 estimated by the regression equations was within 0.0 to 3.0 Kg N/ha of that predicted by DRAINMOD NII. For the
592 CWS cropping system the absolute differences ranged between 0.0 to 9.0 Kg N/ha, and between 0.0 to 3.0 Kg N/ha
593 if the first 5 years of simulation were excluded from the analysis.

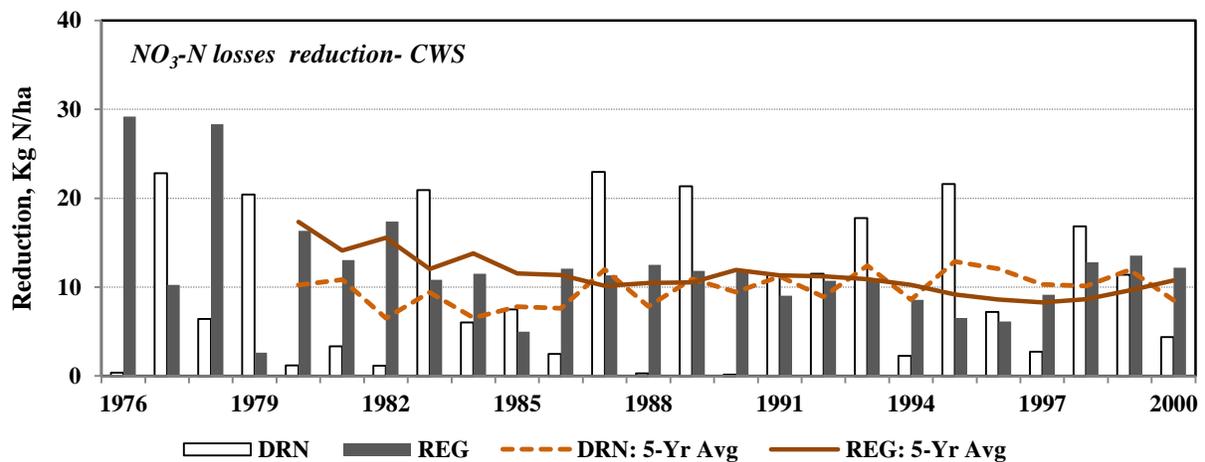


598 Figure 7. Annual NO₃-N losses as simulated by DRAINMOD (DRN) and estimated by the regression models
 599 (REG) for the different management tiers. (Values between parentheses represent average annual drainage
 600 flow ± standard deviation, all units are in Kg N/ha).

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604 **Figure 8. DWM-caused reductions in annual $\text{NO}_3\text{-N}$ losses (Kg N ha^{-1}) as simulated by DRAINMOD (DRN)**
605 **and estimated by the regression models (REG) for continuous corn (CC) and corm-wheat-soybean (CWS)**
606 **management tiers. Solid line represents the 5-year moving average of reduction in annual N-losses (Kg N ha^{-1})**
607 **as estimated by the regression model (REG: 5-Yr Avg), and the dashed line represents the 5-year moving**
608 **average of reduction in annual N-losses (Kg N ha^{-1}) as simulated by the DRAINMOD (DRN: 5-Yr Avg).**

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618 **5 Summary and Conclusion**

619 This article discusses the development of a simple tool suitable for a water quality credit trading system that
620 involves the use of DWM in Eastern North Carolina. In its basic form, the tool is presented by a set of regression
621 equations tracking DWM caused reductions of drainage flows and N mass losses as a function of easy-to-define
622 variables describing site-specific environmental conditions and management practices. DRAINMOD and
623 DRAINMOD-NII were used to conduct long-term simulations of annual drainage flow and NO₃-N losses. Results
624 of the simulations were used to build the regression equations.

625 Simulations comprised numerous scenarios allowing variations in soil type, weather conditions, drainage design,
626 and cultural management practices to represent the many scenarios possible in eastern North Carolina. To
627 accomplish this intensive simulation exercise, efforts were exerted towards data collection and input file preparation,
628 in addition to integrating the DRAINMOD-based models through a batch mode. Regression analysis was performed
629 on the simulated data to develop equations designed to estimate annual drainage rates under four tiers of
630 management scenarios: a continuous corn cropping system operated under conventional drainage (FDCC), a
631 continuous corn cropping system operated under controlled drainage (DWMCC), a corn-wheat-soybean cropping
632 system operated under conventional drainage (FDCWS), and a corn-wheat-soybean cropping system operated under
633 controlled drainage. Similarly, a set of four regression equations developed and assigned to estimate annual NO₃-N
634 leaching losses. The estimations of annual drainage flow and NO₃-N losses by the regression model correlated well
635 with the corresponding DRAINMOD-based simulations ($R^2_{adj} \geq 0.88$).

636 The regression tool was used to estimate annual DWM-induced reductions for 25 years of historical weather data for
637 an experimental site located in Plymouth eastern North Carolina. In comparison to DRAINMOD and DRAINMOD-
638 NII, previously calibrated and validated for this site, the regression models performed well in estimating most of the
639 year-to-year variation in drainage rates and nitrate losses. Exceptions occurred for years with extreme weather
640 conditions or high variations in inter-annual climate conditions; however, the comparison based on a 5-year moving
641 averages illustrated that both methods yielded similar outcomes.

642 In conclusion, this regression approach provides producers, local and federal agencies with an easy-to-use
643 alternative to the burdensome use of process based models to facilitate the assessment of DWM under different site
644 conditions exist in eastern North Carolina. Potential errors with the method can occur when inter-annual variations

645 in rainfall are high which can lead to inaccurate estimations of annual N losses for individual years. To compensate
646 for these errors, the trade of N credits could be based on multiple-year contract. Estimates of N losses average out
647 over five year periods. Although the primary intent of building these regression models was to facilitate the
648 quantifications of N credits, the results presented herein introduce that the tool can be used with different water
649 management applications. Next steps for facilitating wider use of these methods include development of regression
650 equations for other areas where drained agriculture is used such as the Midwest US. Development of a user-friendly
651 interface, web-based tools, and educational program to teach the method will also make the method more acceptable
652 for use in a nutrient credit trading system.

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657 and soil properties.

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